**State Change Manuscript Outline:**

**Introduction**

* Global climate change is forecast to produce dramatic shifts in the frequency and magnitude of extreme climate events across ecosystems worldwide. Increased drought severity, wildfire intensity, and flooding are expected to significantly impact ecosystem structure and function in many contexts.
* To better understand and adapt to these pressures, ecologists have increasingly focused on the concept of resilience – the ability of an ecosystem to resist change and to self-reorganize following a disturbance (Holling 1973, 1986, Gunderson and Holling 2002, Folke 2006).
* Central to the notion of resilience is an emphasis on non-equilibrium dynamics, in which communities or populations exhibit complex trajectories in response to environmental change, such as nonlinear change or hysteresis (Scheffer et al. 2001). Systems which typify non-equilibrium patterns of change are often described as a series of potential discrete “states”. Identification of the state types and the processes which govern movement between them forms a lens by which the effects of environmental stressors may be analyzed (Scheffer et al. 2001, Smith 2011). In this fashion, resilience may refer to the probability that a community type retains its configuration in a given environmental condition.
* Resilience-based frameworks have proven valuable in applied contexts, in which management and restoration are often directed towards maintaining desirable states and coaxing unfavorable states to more desirable ones. This focus on drivers of state change and key thresholds forms a framework that may be used to provide critical interventions when conditions permit.
* Resilience thinking has found purchase in rangeland management, where traditional range models based on successional processes often failed to capture nonlinear community responses.
* Rather than relating rangeland condition to some ideal reference, state and transition models (STMs) describe vegetation change as a series of transitions between discrete community states. In doing so, managers attempt to identify the set of vegetation types that best capture relevant community variation in a system, and the forces that govern turnover between them.
* While a key conceptual tool, classic STMs are based on expert opinion, which may be highly subjective (Allen-Diaz and Bartolome 1998). Empirical tests of STMs, though limited in number, have shown a capacity to overcome biases in state delineation and the enumeration of transition events (Allen-Diaz and Bartolome 1998, Jackson and Bartolome 2002, Bagchi et al. 2012, Stein et al. 2016). Further development of these quantitative STMs is needed to complement expert models by providing clear rationale behind estimates of the number of states, the species by which they are defined, and the probability of transition between them.

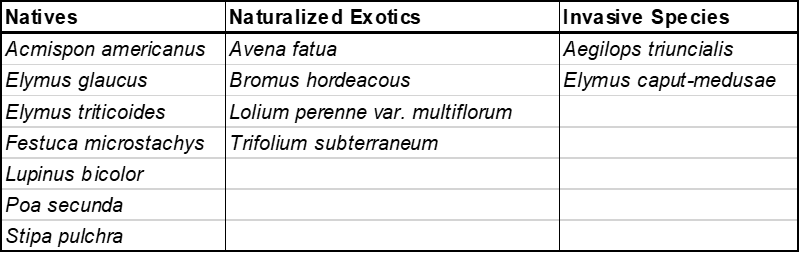
**California Annual Grasslands**

* California grasslands have long been a focal system in the study of nonequilibrium dynamics. Composed primarily of exotic annual species, these grasslands readily shift between dominant groups of taxa (George et al. 1992).
* Conceptually, California grassland vegetation has long been partitioned into several distinct groups, including those based on functional type (grasses, forbs, and legumes) and provenance (native and exotic species). More recently, distinctions are often made between:
  + 1) Naturalized exotic annual grasses that now compose a majority of vegetation in California’s grassland ecosystems.
  + 2) Native perennial grasses and forbs thought to once cover much of the state’s grassland habitat
  + 3) A set of highly invasive annual grasses that are rapidly expanding throughout California rangelands.
* Past work has shown that transitions between annual rangeland types depend on interactions between plant life history strategy, environmental conditions, and contingency in community assembly.
  + Exotic and native grasses exhibit pronounced differences in seed physiology, growth habit, and fecundity. Large-seeded exotic annuals germinate rapidly with the onset of winter rains and invest heavily in reproduction before senescence, resulting in seed rains of over 1,000,000 seeds/m2.
  + Native grasses, on the other hand, often produce far fewer seeds of much lower mass that may fail to compete with annual grasses at early life stages. However, once established, hardy perennials are characterized by low mortality and long-term persistence without active disturbance.
  + Invasive annual grasses exhibit many similar features to naturalized exotics, with a key exception of delayed phenology of growth and seed set.
  + As a result of these differences, seasonal patterns of precipitation and temperature can exert considerable control over productivity and community composition (Pitt and Heady 1978, Reever Morghan et al. 2007). Periodic droughts early in the growing season are thought to produce excess mortality in annual grasses, while late-season rains may favor growth of invasive species.
  + Seasonal patterns of precipitation and temperature can exert considerable control over seedling dynamics, productivity, and propagule production. Periodic winter droughts may reduce the abundance of rapidly germinating annual grasses, while consistent precipitation (particularly late in the growing season) may favor growth of invasive species with later phenology.
  + Native perennial grasses are thought to be highly recruitment limited, failing to produce seed rains comparable to those of their annual counterparts, though are strong competitors once established.
  + Exotic annuals (and invasive grasses, in particular) appear to exhibit strong priority effects through changes in nutrient cycling and deposition of thick litter layers that impede competitor establishment.
* Current climate projections emphasize increased duration and intensity of drought events in California which may act as critical tipping points in many ecosystems (Shaw et al. 2009, Pierce et al. 2012) – a recent drought of historic magnitude has been shown to divide many species into distinct sets of “winners” and “losers”, with dramatic changes in productivity and biodiversity at the community level (Prugh et al. 2018). With pressing concerns on the ability of these communities to resist and reorganize following disturbance, quantitatively-driven development of state and transition models in California annual grasslands may provide a key evaluation of resilience in a system heavily impacted by extreme climate events and species invasion. Indeed, many global grasslands mirror these observed trends of dominance by annual invaders and increasing climatic fluctuations.
* Here, we assess interactions between community assembly and climatic variation on vegetation composition in California annual grasslands across a 10-year period encompassing recent drought (2013-2016) and potential recovery. Using data from experimental plantings of three key grassland species groups – naturalized annual, native, and invasive species – we aim to test key assumptions of the resilience of different communities and the potential drivers of transition between them.
  + What states best partition observed variance in plant community composition? What species define these states?
  + What states arise with different planting composition mixtures? As these states are observed over time, are transitions between states characterized by continuous, reversible changes or non-reversible changes?
  + How do key drivers of community composition (assembly order and climate) govern transitions between states?

**Materials and Methods:**

Study site

* Field plantings were conducted in research fields at the University of California, Davis (38.545751, -121.784780). Soil information, land use history, etc.
* Prior to planting, soil was disked, irrigated, and received a broad-spectrum herbicide (glyphosate) to remove the existing seed bank.
* Three planting mixtures were established based on existing state-change models of California grassland systems, or common delineation between community types (Table 1). For all possible 1-, 2-, and 3-group planting combinations, we established eight 1.5m x 1.5m plots (2.25 m2; 56 plots total).
* What is the detail for planting amount and number of seeds added?
* In each growing season from 2008 – 2018, total areal cover of all species was estimated visually to the nearest 10%. Cover observations for each species were performed in mid-and late-spring to capture the season of maximum percent cover for each species, because any one sampling wouldn't account for variation in species phenology.

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Weather data

* Weather data was provided by a local California Irrigation Management Information System (CIMIS) monitoring station in Davis, CA (38.535694, -121.777636). CIMIS automated dataloggers collect weather data on a minute-by-minute basis, including air temperature, soil temperature, precipitation, solar radiation, vapor pressure, and wind speed. We aggregated these data into monthly intervals, where we calculated Standardized Precipitation-Evapotranspiration Index (SPEI), a metric of drought stress (*D­i*) at a given timepoint, *i*:



* Where *Pi*represents observed precipitation and *ETo­i*represents estimated evapotransporation. *ETo* was calculated using the Penman-Monteith equation, defined as:



* Where *Rn* is net radiation, *G* is soil heat flux, *(es – ea)* isthe vapor pressure deficit of air, *ρi* is the mean air density at constant pressure, *cp* is the specific heat of air, Δ is the slope of the saturation vapor pressure temperature relationship, γ is the psychometric constant, and *rs*and *ra* are the surface and aerodynamic resistances (FAO).
* SPEI offers flexible, variable timescale estimations of drought stress that can be used to quantify the effects of multi-year climate patterns (Vicente-Serrano et al. 2010). For each year between 1980 and 2018, we calculated SPEI for a single water year (October – May; 8 months), two consecutive water years (20 months), and three consecutive water years (32 months). We then standardized these values by fitting the drought index series to a log-logistic distribution. All SPEI calculations were performed using the package “spei”.

Delineation of States

* Due to intermittent invasions by agricultural weeds, community analyses were performed on a subset of the total community corresponding to species that were planted in our initial mixes, in addition to *Bromus diandrus*, a common naturalized annual grass. The resulting dataset captured 93% of the total vegetation abundance observed over the course of the experiment.
* Per Allen-Diaz and Bartolome (1997), quantitative generation of State-Transition models can be performed by algorithmic partitioning observed variance in vegetation composition. To this end, we chose to apply an unsupervised clustering algorithm (k-medioids clustering) across the total variation in community composition observed within our dataset. K-medoids clustering randomly selects *k* of *n* total datapoints as group “medoids” and computes the sum of distances between points and their associated medioid, based on Bray-Curtis dissimilarity. This algorithm then iteratively swaps these mediods and recalculates summed distance to achieve a solution that best captures the total variance of the data. R library used – “pam”.
* To determine the most appropriate number of states, we applied k-medioids clustering across values of *k* from 2-10. We then subjected the output of each of these runs to a battery of tests (list tests here, if needed); the value of *k* with the most consist performance across all tests was used to determine the number of clusters that best represented discrete partitions within this dataset. R library used – nbclust.
* Following the partition of states, we then conducted indicator species analysis to establish what species are associated with each state. Indicator species analysis was conducted using the “vegan” package.

****Construction of State-Transition Models

* Following the association of observations to discrete states, we fit a multistate model (aka Markov model) to the data. Multistate models represent systems where subjects transition between a set of discrete classes over time and may be uniquely suited to examining state and transition models through a statistical framework.
* In our analysis, we constructed a multistate model consisting of all states identified in clustering analysis, with probabilities fit to all possible transitions between states.
* To test for effects of initial planting composition and climatic variation on the probability of state transition and resilience, we added a series of covariates to multistate models that correspond to SPEI and the presence of state indicator species in the initial planting composition.
  + ****E.g, the probability of a transition between states 1 and 2, *q’12*, can be represented by:

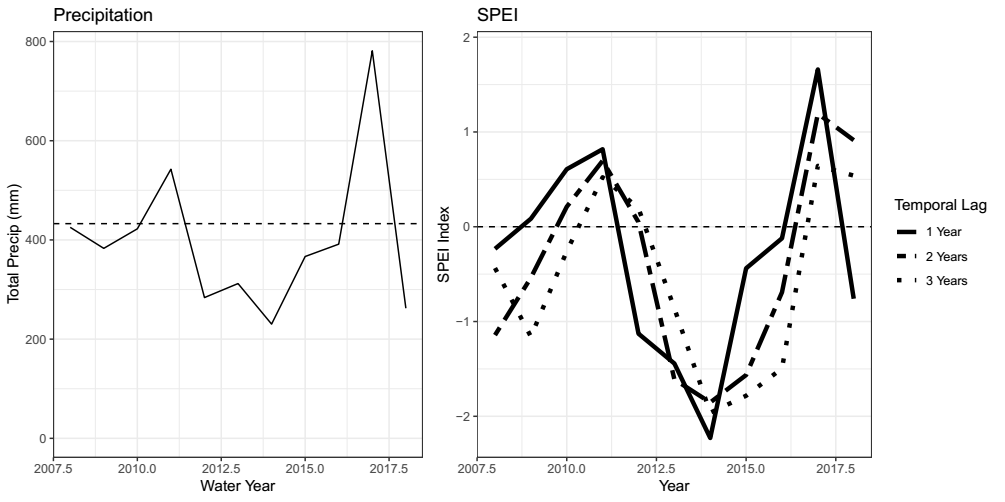
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* + Where q12 is the baseline probability of a transition, and β1 and β2 are coefficients fit to recorded SPEI values and planting composition, respectively.
* After fitting models with and without SPEI and initial planting covariates for 1-, 2-, and 3-year drought indices, we then calculated AIC scores for each model. We selected the model with the lowest AIC score (ΔAIC < -2) as our best fit model. Further comparisons between subset models containing nested sets of parameters were made using likelihood ratio tests.
* Multistate model fitting and model selection was performed using the “msm” package.

**Results**

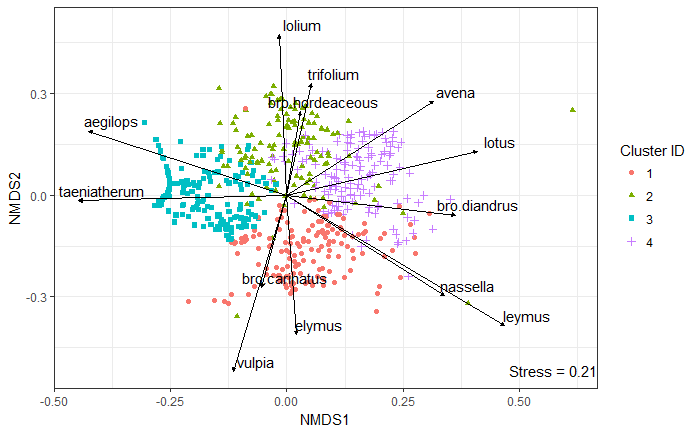
1. **SPEI Figure**

* I suppose this could also be in the methods part of the paper.
* 1st panel shows mean annual precipitation in mm, with average from 1983 – 2018 as a dashed line
* 2nd panel shows the value of the SPEI index over the course of our study, standardized relative to climate patterns from 1983 – 2018. These values are roughly scaled between -2 and 2, where a value of -2 is a historic drought, and a value of +2 is a historic wet period.
* Key takeaway here is that there was a really significant, measurable drought in 2012-2016. This isn’t surprising, but helpful to demonstrate that this metric is sensitive to these patterns, and does a good job of capturing the drought.
* Also interesting to note how the intensity of the drought varies depending on what sort of spatial lag we’re looking at. If we have a cumulative sum over 3 water years, for example, the drought extends all the way into 2017, but 2016 seems like a relatively normal water year.
* 2018 data seems surprisingly low. Is this a true value? Need to double-check. Also, curious why the 1 year temporal lag does not show the exact same pattern as the raw precip data. A couple possible reasons – the first is that the log-logistic data transformation standardizes the data in a way that skews the trend a little, the other is that focusing on November – May precip in the second panel ignores some early and late season precipitation that is captured in the first.



1. **NMDS of state assignments**

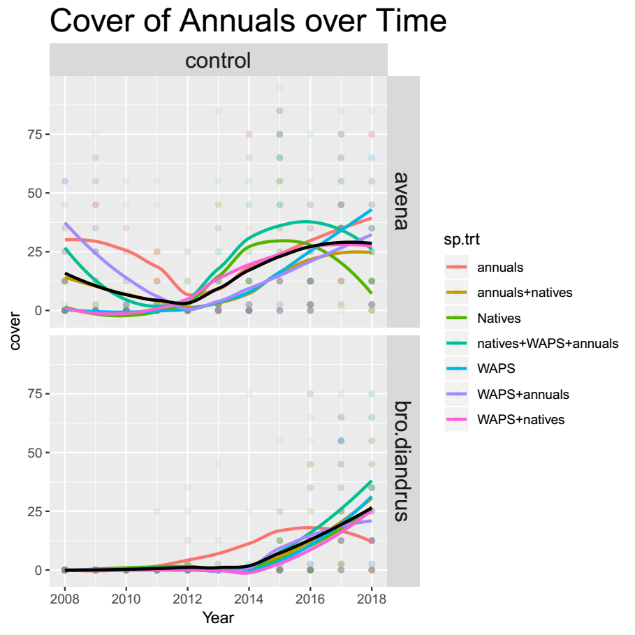
* Still need to fix the species labels to reflect modern taxonomy
* NMDS figure shows total variation in community composition for all observations between 2008 – 2018.
* This is not a particularly interesting figure itself, but represents the total amount of variation seen among communities over the course of sampling. **Makes more sense when accompanied by the next figure (indic. species analysis table)**



1. **Indicator species analysis table**

* As before, still need to fix the species labels on this
* Key point is that there seem to be four separate groups here, 3 of which that loosely fall along the native / invasive / exotic annual lines, but there seems to be a fourth group that is primarily annual exotics that are strong drought tolerators.

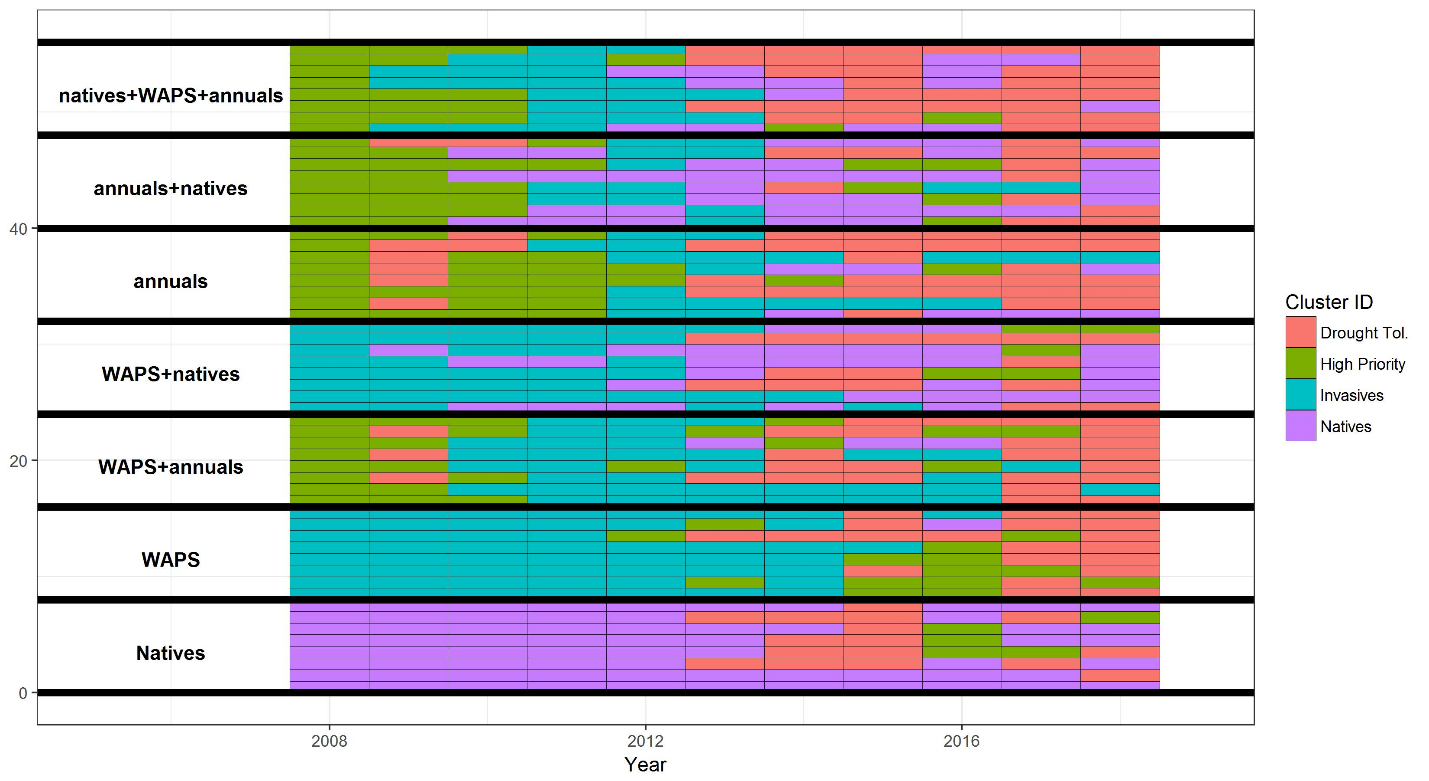


Per discussion of differences in Avena and Bromus Diandrus abundance, I’ve included a figure of the relative abundances of these two species. I don’t see a clear pattern here of Avena responding better post-drought than during the drought itself – in fact, it seems that Avena grows in relative abundance starting around 2014, then saturates in a couple years. This corresponds fairly closely to the prevalence of the “drought tolerators” state, which really only appears to predominate starting in 2014.

It’s possible that looking at relative abundances makes this result conflict with your intuition from the field. I’m not sure if there’s any easy way to get around this, as the metrics I am using rely on compositional similarity, which don’t capture differences in total abundance.

1. **Group assignments by individual plot**

* Will need to change the labels on the y-axis for this to reflect terminology used in methods section. Can also make the width of the bars smaller.
* This is a busy graph, but I think it’s important to highlight what states first arise when you vary planting composition, in part because the pattern is very clear.
  + Adding annuals to any planting mixture makes them a dominant part of the community in the first year
  + WAPS appear later on in all treatments where they are added, in addition to annuals (strong native resistance)
  + Natives rarely do well early on, but becoming bigger parts of the community later on.
* Suggestions on ways to improve readability of this figure?

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1. **Transition assignments over time and transition frequency table**

* Transitions appear to be frequent and widely distributed in this dataset. All possible transitions occurred, but some appear more common than others – natives appear to have rarely transitioned to groups 2 or 3, for example.



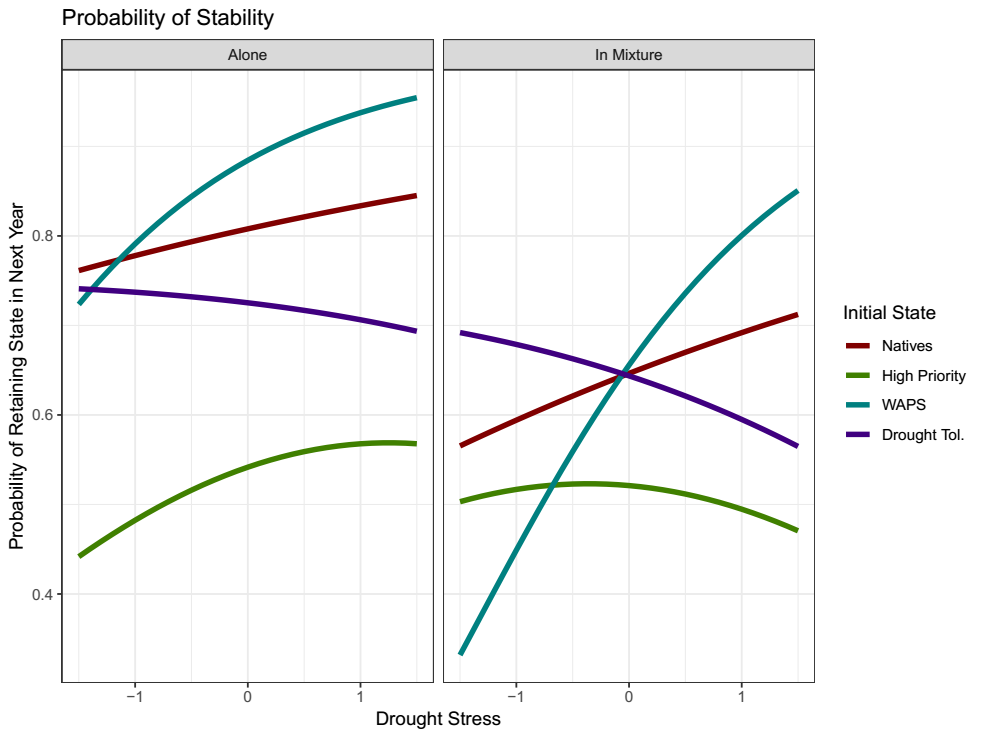
1. **AIC model selection table**

* Results of Markov model fitting and model selection. Compares models with the raw transition matrix frequencies, planting composition covariate, and multi-year drought covariates.
* Shows that the best fit model in this case (lowest AIC score) is one that contains covariates related to planting composition and 1 year drought. The 3 year drought model is also an acceptable alternative, and the 2 year drought model isn’t a whole lot worse than the other two.



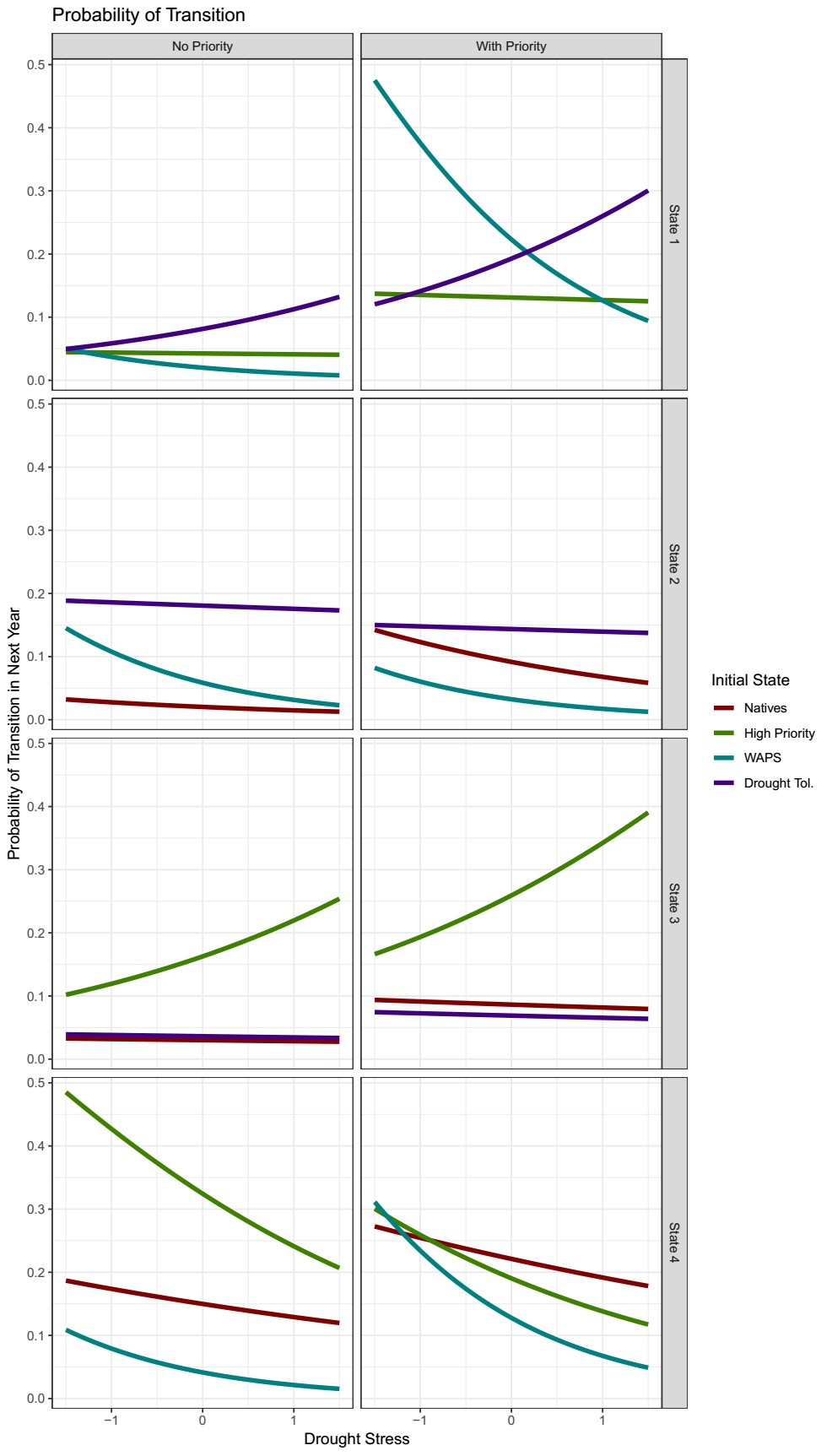
1. **Stability probabilities**

* This figure shows the probability that a community with a given state assignment retains that assignment in the next year. Quantification of resilience.
* Generally, the pattern I see here is that:
  + The high priority state doesn’t stick around very long, regardless of drought stress
  + WAPS are very common in wet years, seem to do more poorly when conditions are dry
  + Natives are quite stable, but don’t show as strong of a drought response as other groups. They’re somewhat more stable when conditions are wetter.
  + Drought tolerators, unsurprisingly, prefer drought conditions.
  + In all cases, more speciose planting mixtures decrease the odds of retaining your state over time.



1. **Transition probabilities**

* The following figure shows the probability that a community in a given state transitions to a focal state in the next year. In this case, we have the focal states as rows, with odds of transition from previous states as colored lines. The columns show a split between whether that focal state was or was not represented in the initial planting. The states, in order, are natives, high priority, invasives, and drought tolerators.
* I think there are a few ways to clean up this figure to make it easier to interpret. I’d like to label each row with “Probability of Transition to Natives”, “Probability of Transition to Invasives”, etc.
  + Stein et al. split this same figure up into 12 separate graphs, with pairwise transitions between two states shown. However, this can be quite messy depending on the number of parameters in the model (they had just 1 quantitative variable, RDM).
  + Any other thoughts?
* My key takeaways:
  + Transitioning to native states is possible for all communities at some point, but only if they were able to colonize early on.
  + As before, communities often transition to WAPS when it’s wet, Drought tolerators when its dry.
  + While high priority species are well represented early on, it seems difficult for them to reappear when things are left mostly static.

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**Discussion**

*Revisiting previous questions:*

1. What states best partition observed variance in plant community composition? What species define these states?
   1. It seems that communities roughly follow the fluctuations in abundance predicted by the standard 3-state perspective of California annual grasslands (with the caveat being we planted them in this way!), with one key exception – there seem to be a subset of these species that do particularly well during and post drought that are distinct from the species that do well early on.
2. Are transitions between states characterized by continuous, reversible changes or non-reversible changes?
   1. We find that many transitions between communities are reversible and highly frequent, corresponding with our notion of these communities as being dominated by non-equilibrium dynamics.
   2. However, over the course of our experiment, we also found that certain states varied considerably in their resilience and transition direction. Native states appeared to be particularly stable, while others, such as high priority annual grasses, dominated many planting compositions early on, but were not particularly stable.
3. How do key drivers of community composition (assembly order and climate) govern transitions between states?
   1. Consistent with reported invasions of exotic annual grasses, transitions between annual dominated states and invasive species are frequent, particularly in wet years.
   2. However, severe drought appears to have dramatic effects on both the stability and persistence of different states. States dominated by invasive species, which exhibit later phenology, were likely to shift to a more drought-tolerant state when during the historic drought from 2014 – 2016.
   3. Assembly order continued to have large effects on patterns of community turnover, years after planting.
      1. This effect was most pronounced in “native” states – even when native species were not dominant early on, planting compositions that contained native species were more likely to transition to a native state later on. Conversely, communities that did not receive any native seed very rarely experienced a state transition.
         * Important for restoration; native species may appear when conditions are favorable, even if not abundant early on. Consistent with notions of spatial storage effects in perennial grasslands.
4. Some important caveats
   1. State-transition approaches are great at distilling down temporal community dynamics into groups of species that have similar patterns of abundance, but may not yield particularly nuanced insights.
      1. While interpretation of state-transition models can be informed by other studies of individual species dynamics, it’s difficult to determine within-state differences in species abundance in state change models. In our case, this may be an inability to determine whether species are responding to drought, or immediately afterword.
   2. This sort of analysis is limited to the scope of total community variation observed within a given time series of observations – quantitative analysis can be used to complement and test predictions made by expert models, not necessarily to create new models from scratch.
      1. Detection of environmental parameter effects depends on variation observed (what will happen with drought recovery, for example?)
      2. Limited to the species pool present in a given site. If state-transition models are meant to describe the phases and states that may appear within a given soil type or management context, these all need to be present to be tested in a quantitative fashion.

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**Supplemental Information**

1. **NBClust k selection test output**



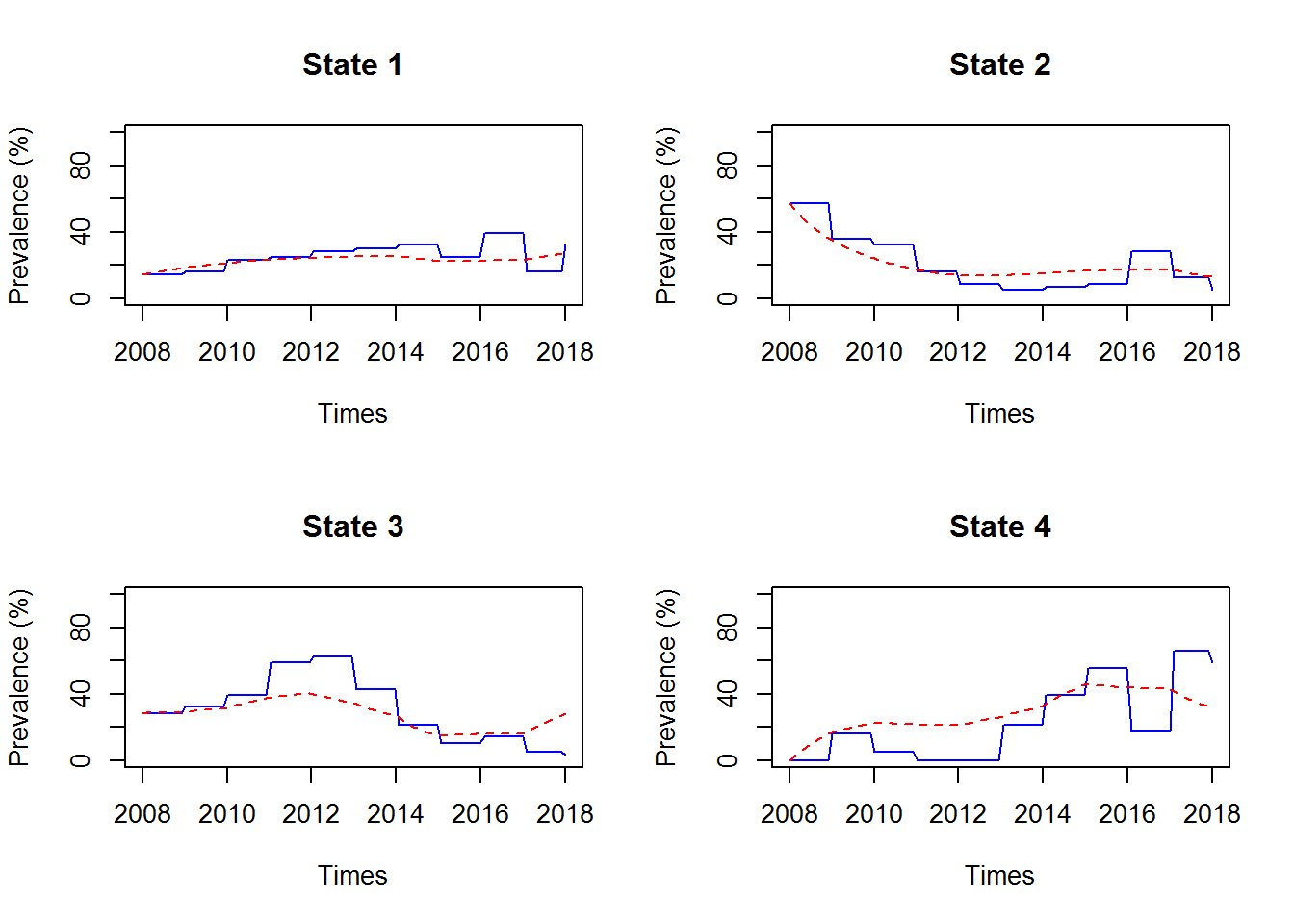
1. **Visualization of the relative percent cover of all species by state assignment**

* Older figure, need to think of a better way to convey this information, if needed.



1. **MSM model output, estimated coefficients, visualized fits**

* Can also include the coefficient table, hazard ratios, etc.

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